



Environmental Engineering

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Removal of microcystin-LR from water by ozonation

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ABSTRACT: In temperate climates, cyanobacteria (blue-green algae) occur most frequently in summer, when the demand for recreational water is highest. Blue-green algae can produce toxins and are also responsible for the taste and odour of water, which significantly impairs its quality. Cyanotoxins are very dangerous substances which can intoxicate hepatocytes and the nervous system in humans and animals. To prevent such situations, it is very important to remove cyanotoxins from water effectively during the pretreatment process.

In the present study, the ozonation of water containing microcystin-LR was tested. We performed this research at the laboratory scale as well as in a pretreatment plant near the Sulejow artificial lake during several seasons.

Keywords: Cyanobacterial toxins, microcystin, ozonation, water pre-treatment.

1 INTRODUCTION

Cyanobacteria (blue-green algae) are organisms that have some characteristics of bacteria and some of algae. They are similar to algae in size and, unlike other bacteria, they contain blue-green and green pigments and can perform photosynthesis. Pollution related to human activities, e.g. from agricultural runoff and inadequate sewage treatment, has led to excessive eutrophication (fertilization) of many water bodies. As a result, there is excessive proliferation of cyanobacteria in fresh water which has a considerable impact upon recreational water quality (as they are responsible for the taste and odour of water). In temperate climates, cyanobacterial dominance is most pronounced during the summer months, which coincides with the period when the demand for recreational water is the highest. The formation of cyanobacterial blooms is not a new phenomenon. The earliest reliable account of such blooms was at the end of the twelfth century (Ressom et al. 1994).

Blue-green algae produce several toxins, including neurotoxins, hepatotoxins, cylindrospermopsin and lipopolysaccharide endotoxins (Carmichael 1997). Toxic cyanobacteria are cosmopolitan. They have been recorded on every continent and about 50–75% of tested cyanobacterial blooms have been toxic (Codd 1995). Cyanobacteria have been implicated in various episodes of human and animal illnesses in Europe (Turner et al., 1990), North and South America (Billings 1981, Jochimsen et al. 1998), Asia (Yu 1989, Ueno et al. 1996), Africa (Zillberg 1966) and Australia (Falconer et al. 1983).

The type of cyanobacterial toxins most frequently found in fresh waters is microcystin-LR. This is a hepatotoxic peptide produced by a number of cyanobacterial genera, the most notable of which is the widespread

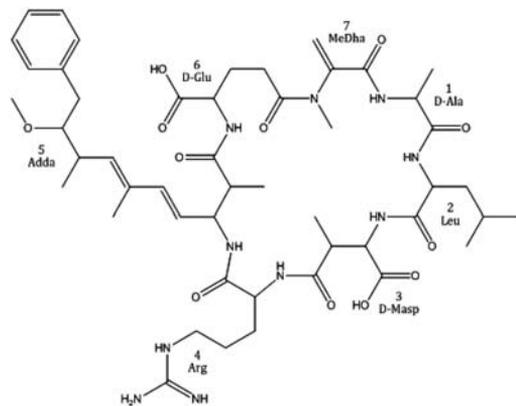


Figure 1. The structure of microcystin-LR.

Microcystis, from which the toxins take their name. The microcystin-LR is a cyclic heptapeptide (Figure 1). The molecule consists of a seven-membered peptide ring which is made up of five non-protein amino acids and two protein amino acids: Leucine (L) and Arginine (R) in positions 2 and 4. The Adda side chain (position 5) is a key structural element necessary for biological activity. Separation of the Adda component from the cyclic peptide renders both components non-toxic (Carmichael 1992).

Many strategies for the removal of microcystin-LR from water have been investigated. Conventional methods of water treatment (sedimentation, filtration, coagulation) have been reported to be ineffective (Hoffman 1976, Himberg et al. 1989).

Activated carbon can remove microcystin-LR, but doses have to be higher than those generally used in water treatment (Falconer et al. 1989). Nicholson

et al. (1994) have reported that microcystin-LR can be removed by chlorination, but unfortunately chlorination of organic compounds can also cause adverse health effects in humans. Chlorine dioxide is not effective at those doses used in drinking water treatment. Hydrogen peroxide was found to be ineffective in toxin removal; used alone or with UV radiation, it can remove only about 50% of microcystin-LR after 30 minutes (Rositano & Nicholson 1994). Ozone has been found to be most effective in oxidation of cell-bound microcystin, if it is applied at a sufficiently high dose and with a sufficiently long contact time. Dissolved air flotation has been proposed in which the recycled water is saturated with ozone-rich air (Baron et al. 1997). Ozone-rich air has also been proposed for dispersed air flotation. These approaches might result in the reduction of extracellular toxins as well as enhanced cell removal (Chorus & Bartram 1999).

This paper presents the results obtained by the ozonation of microcystin-LR in a water treatment plant near the Sulejow reservoir in Poland and at the laboratory scale. The studies concerned with the processes of water production were conducted during three seasons with different ozone concentrations. This is the first research work on such a scale.

2 MATERIALS AND METHODS

To remove mechanical, organic and inorganic pollutants, the water samples were first filtered using a membrane and then concentrated using the solid phase extraction method. After sample pre-concentration, microcolumns were rinsed with methanol. The alcohol fraction was evaporated at inert gas flow at room temperature. After evaporation the samples were dissolved in acetonitrile–ammonium acetate buffer (74:26 v/v) filtered through a 0.45 mm filter and separated by RP–HPLC (Meriluoto 1997).

The concentration of ozone was determined by iodometric titration (Rakness et al. 1996).

3 RESULTS AND DISCUSSION

The changes in microcystin-LR concentrations (from an initial concentration of 16 mg/dm³) during the ozonation process at the laboratory scale can be found in Table 1.

The total toxin content was oxidized after the addition of about 0.01 mg O₃ per 1 µg of microcystin-LR. This result suggests that about 220 ozone molecules combine with one molecule of microcystin-LR. It is highly probable that the peptide ring of the toxin molecule is broken, but the mechanism has not yet been identified. The changes in microcystin-LR concentrations in relation to different ozone levels are given in Figure 2.

The amount of oxidizing microcystin-LR per unit of time, along with the given speed of ozone addition, does not correlate with the microcystin-LR

Table 1. Changes in microcystin-LR concentrations (from an initial concentration of 16 mg/dm³) during the ozonation process (laboratory scale).

Time of ozonation [s]	Ozone concentration in sample [mg/dm ³]	Mean yield of microcystin-LR residue [%]
0	0.00	100
5	0.06	78
10	0.11	38
11	0.12	32
12	0.13	25
13	0.14	17
14	0.15	11
15	0.16	5
16	0.17	0

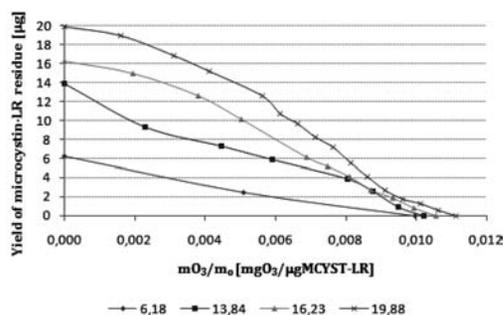


Figure 2. Changes in microcystin-LR concentrations in relation to ozone levels for different concentrations.

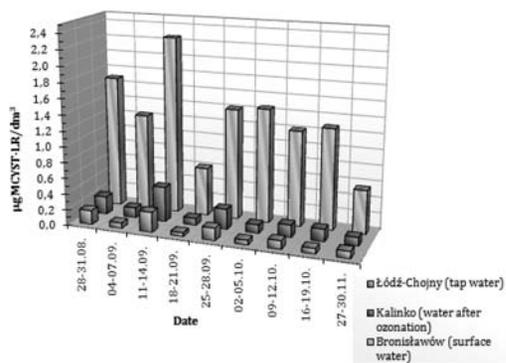


Figure 3. Concentrations of microcystin-LR in surface water, after ozonation in the Kalinko treatment plant, and in the Lodz-Chojny pumping station in the 2001 season.

concentrations in the solution. The microcystin-LR ozonation process is useful when cyanobacterial cells are removed.

The concentrations of microcystin-LR in surface water from the Sulejow artificial lake, after ozonation in a water treatment plant in Kalinko and at a final step in the Lodz-Chojny pumping station, are given in Figures 3, 4 and 5.

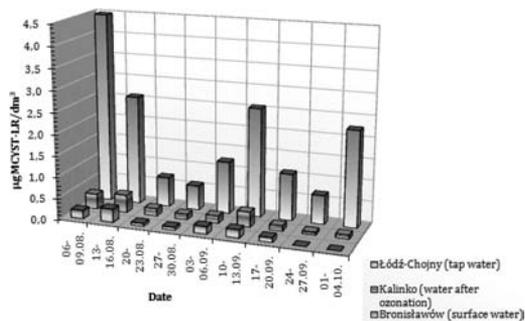


Figure 4. The concentrations of microcystin-LR in surface water, after ozonation in the Kalinko treatment plant, and in the Lodz-Chojny pumping station in the 2002 season.

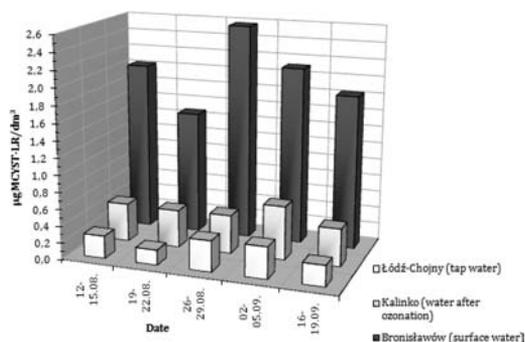


Figure 5. Concentrations of microcystin-LR in surface water, after ozonation in the Kalinko treatment plant, and in the Lodz-Chojny pumping station in the 2003 season.

Table 2. Doses of ozone used in the treatment system in Kalinko.

Season	Ozone doses [g/m^3]		Mean yield of ozonation [%]
	Min.	Max.	
2001	1.5	3.0	90.8
2002	1.4	3.5	92.1
2003	1.3	1.7	85.7

The ozone doses used in the treatment system in the Kalinko station can be found in Table 2.

Toxic cyanobacterial blooms were observed between 2001 and 2003. Microcystin-LR was found in surface water samples from the Sulejow reservoir even two months after the blooms had decayed.

4 CONCLUSIONS

The final efficiency of microcystin-LR removal in the ozonation process ranges between 82% and 100%; this results in concentrations lower than the maximum recommended by WHO ($1 \mu\text{g}/\text{dm}^3$). This efficiency can

be obtained using ozone doses of $1.3\text{--}3.5 \text{ g O}_3/\text{m}^3$ and a contact time of 35–48 mins.

After the ozonation process, water did not exhibit toxic properties. In spite of that, the products formed in the ozonolysis reaction may have adverse health effects. The toxicity of such substances is still under investigation.

Using a combination of ozone and UV light is possible. This method may enhance the efficiency of microcystin-LR removal but it has some disadvantages, e.g. it is a highly energy-consuming process. The application of such a technique may result in a higher cost for water production.

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